

# Methyl Mercury Exposure through Seafood Diet and Its Effect on Aquatic Life and Health in United Arab Emirates

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## Abstract

Methylmercury bioaccumulates and concentrates through the aquatic food chain and is known for its potent toxicity, exposure to methylmercury is linked to increased health risks of humans as well as aquatic life. This paper reviews the existing comprehensive literature highlighting the interaction between Dissolved organic matter (DOM) and MeHg, the process of bioaccumulation and current unresolved risks due to MeHg contamination in aquatic life and its related health hazards. This review particularly focuses on the dietary exposure to methylmercury through seafood in UAE. Though further research is warranted on new factors that influence MeHg uptake and toxicity and bio-indicators of exposure to MeHg, that leads to cardiovascular and neurologic effects in adult populations in UAE, it has been concluded that to protect human health and aquatic life, it is important to take necessary steps in order to eliminate or at least reduce the sources of methylmercury dietary exposure to national population.

**Keywords:** methylmercury, contamination of aquatic life, UAE, seafood risks

## 1. Introduction

The United Arab Emirates (UAE) has rapidly grown in the last 40 years from a subsistence and nomadic fishing population of about 400,000 to a diversified multicultural and industrial based population of more than 4.4 million (Cassen, 1978; United Nations, 2007). The modernization in UAE has been more rapid than any other nation in the history of the world. Two of the cities Dubai and Abu Dhabi in UAE are well known international urban centers and by some measures Abu Dhabi is one of the world's richest cities (Gimbel, 2007). UAE owns approximately 8 percent of the total world's oil reserves which is contained in Abu Dhabi, the largest emirate of UAE out of the seven emirates (U.S. Energy Information Administration, 2010). This rapid development can be mainly attributed to oil exports, through seafood, mainly shrimps and fish are also of significant value for both local consumption as well as export revenue in Arabian Gulf area. And therefore, for various socio-economic reasons it becomes imperative and crucial to maintain good and healthy aquatic environmental quality (Price et al., 1993). In UAE, there has been a significant improvement in the past few decades in national health due to the diversified food supply. Life expectancy (at birth) increased from 53 years in 1960 for UAE nationals, to 65 years in 1975, and then to 78 years in 2006 (WHO, 2007). Apart from increase in life expectancy due to the technology revolution, the new dietary pattern has posed various health risks mainly consist of growing dietary exposure to anthropogenic contaminants. In addition, the recent data also revealed that the country is also recorded the growing prevalence of obesity (Davidson, Krometis, Al-Harthi, & Gibson, 2012).

Over the past few decades marine contamination has been a leading concern in the wake of rapid urbanization and industrialization of Arabian Gulf coastal areas. Metal contamination is of particular concern as heavy metals such as cadmium, mercury and zinc are frequently discharged resulting from shipping traffic of Gulf or from municipal and industrial effluents, and these metal contaminants can enter through the food chain and can bio-accumulate in the tissues of living organisms (Fleming et al., 2006). Major concern is mercury contamination which causes pollution of seafood such as shellfish, fish, oyster and other variety of seafood especially in densely populated places (Bortoli, Gerotto, Marchiori, Palonta, & Tronco, 1995). Most scientifically challenging contaminants out of the other metal contaminants that threatens the aquatic resources are the increasing levels of mercury (Hg) (Krabbenhof, 2003). Methyl mercury (MeHg) bio-accumulates in fish, a species that is at the top-level of aquatic food chain, MeHg is an organic component that conforms from

inorganic mercury and its concentration in fish is approximately 1-10 million times higher than MeHg concentrations found usually in surrounding waters (US Environmental Protection Agency USEPA, 2001; Lawrence & Mason, 2001). The mercury toxicity is mainly transported by dissolved organic matter (DOM) (Strober et al., 1995) which is the major contributor to the pool of organic matter in sea water. Thus, it plays a vital role in the global carbon cycle (Mopper & Degens, 1979; Cauwet, 1978). Further, DOM strongly interacts with mercury, thereby affecting its mobility, speciation, toxicity and solubility in the aquatic environment (Ravichandran, 2004). This paper reviews the process of bioaccumulation of methylmercury through DOM and its effect on the aquatic life and the entire food chain; the study also presents a broad risk assessment of national public health of UAE due to dietary exposure to MeHg.

## 2. Method

Standard literature review method has been used to evaluate the effect of methylmercury dietary exposure through seafood in UAE. Various empirical studies and journals were reviewed to evaluate the process of bioaccumulation of methylmercury in aquatic environment through DOM and its subsequent effect on the entire food chain. UAE-specific studies were reviewed to examine the dietary exposure to methylmercury and its effect amongst the national population.

## 3. Results

The review of relevant studies have been categorized into three sections, the first section shows what is DOM and how it interacts with methylmercury in aquatic environment. The next section describes the associated risk of methylmercury in aquatic environment and on the entire food chain and last section provides an insight into formation of methylmercury, its dietary exposure in UAE and its associated risks.

### 3.1 Dissolved Organic Matter

DOM (Dissolved Organic Matter) is composed of various organic soluble materials that are derived from partially decomposed organic materials (e.g. plant residues and soil organic matter) and other soluble particles released by living organisms (e.g. algae, bacteria, and plants). It is present in all soils dissolved either in flood water or soil solution, but its concentration is usually heavy in aquatic ecosystems and wetland than in agricultural soils. Thus it plays a significant role as it is a source of carbon for microorganisms and also helps in nutrient supply and sequestration (Wright & Reddy, 2009). DOM acts as an organic substrate that fuels the growth of heterotrophic bacteria and algae (Tuchman, 1996; Bernhardt & Likens, 2002). The DOM composition in water body is based on thousands of molecules (Sleighter et al., 2010; Stenson et al., 2003). The nature of source materials and the carbon cycling processes within the ecosystem, that includes terrestrial watershed influences dissolved organic matter's chemical characteristics in a water body (McKnight & Aiken, 1998; Aiken & Cotsaris, 1995; Fellman et al., 2010). Hence the origin, chemical composition and fate of the various DOM pool components in aquatic systems are not known entirely (Hansell, 2002).

As for its molecular composition DOM has a complex mix of molecules and the unique presence of these molecules in DOM reacts with mercury (Hg) varies. For example, studies of DOM binding mercury have shown that only limited fragment of molecules present in DOM possess the essential functional groups that can help in strongly binding mercury (Haitzer et al., 2002). Even though small fragments of DOM molecules react with Hg, usually large fractions of molecules in DOM are inert with respect to interactions with Hg, particularly under conditions where the concentration of mercury (nanograms per liter) is much less than DOM (milligrams per liter) (Gerbig, Ryan, & Aiken, 2012).

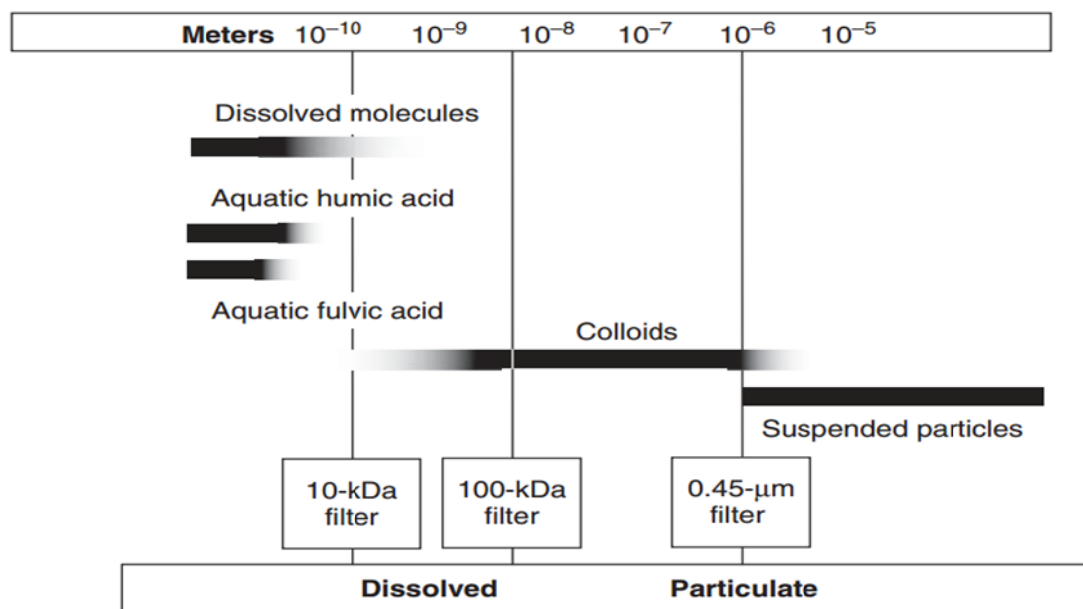


Figure 1. Conceptual diagrammatical representation the distribution of organic matter by size between dissolved organic carbon and particulate in natural waters (Gerbig, Ryan, & Aiken, 2012)

Basically DOM comprises a complex mix of molecules; with heterogeneous continuum of low to high weight of molecules of organic compounds that exhibits different reactivity and solubility. As shown in Figure 1 organic matter are arbitrarily divided into particulate and dissolved organic carbon in natural waters, this division is based on filtration process, which happens generally via 0.45- or 0.7-µm filters (Gerbig, Ryan, & Aiken, 2012). The situation further complicated, in certain processes, of compound pool that DOM is comprised of which may counteract with each other with respect to particular reaction/reactions. The presence of both the compounds (dissolution- inhibiting and dissolution- promoting) is responsible for DOM interactions with bulk and colloidal nano-particulate HgS (s) (Waples et al., 2005; Slowey, 2010). The significant factor that contributes and controls DOM reactivity with mercury are the differences in carbon content and variance in the nature and amount of DOM present in the ecosystems.

### 3.1.1 Interaction between Dissolved Organic Matter (DOM) and Methylmercury

More than 60 percent of dissolved Hg in coastal waters is associated with suspended particles and organic matters (Fitzgerald & Lyons, 1973). Perhaps one of the key aspects of the cycle of mercury in a water body is its affinity with ligands that are presented in DOM (Driscoll et al., 1995; Lamborg et al., 2004). Dissolved Organic Matter in aquatic environment is ubiquitous and is bound to trace metals; in fact there is increasing evidence that the interaction of DOM with mercury affect their solubility, speciation, toxicity and mobility (Buffle et al., 1988; Loux, 1998). Several previous studies have also shown that there is heavy interactions between DOM and Hg and have indicated that there is a positive correlation between the concentrations of DOM and Hg in many natural waters (Lindberg & Harriss, 1974; Andren & Harriss, 1975; Meili et al., 1991; Mierle & Ingram, 1991; Hurley et al., 1995; Driscoll et al., 1995; Watras et al., 1995; Kolka et al., 1999; Baeyens et al., 1996; Shanley et al., 2002).

The effect of DOM depends on the composition of the micro flora on the uptake ionic mercury  $Hg^{2+}$  by microbes as bacteria that reduce iron and sulphate can transform mercury into neurotoxic MeHg which promptly biomagnifies in the food web. On the contrary, complexes of  $Hg^{2+}$  can restrict the methylation process, thereby helping to reduce the toxicity and bioaccumulation of this metal (Moye et al., 2002; Drexel et al., 2002; Haitzer et al., 2003; Gorski et al., 2008).

Though it is clearly understood that MeHg is biomagnified and bio-accumulated in aquatic food chain, the environmental factors that influence or control the process of methylation of mercury is not clear. Studies have shown that even low concentrations ( $\sim 3$  mg/L) of natural DOM in aquatic environment influence the methylation and biological uptake of Hg by strongly complexing the  $Hg^{2+}$  and methylmercury (MeHg) with organic and inorganic ligands presented in DOM (Dong, Liang, Brooks, Southworth, & Gu, 2010). On the other hand, study

by (Ravichandran, 2004), shows that DOM affects the production as well as bioaccumulation of MeHg, fish bioaccumulates mercury the most. The photochemical reduction of  $\text{Hg}^{2+}$  to elemental mercury DOM plays a key role and subsequently re-oxidation of elemental mercury to  $\text{Hg}^{2+}$  as a result of this process; it affects the bioavailability and volatilization loss of mercury to organisms. Studies by researchers proves that strong ligands presented in high molecular weight (HMW) DOM can effectively bind Hg thus reducing the availability of  $\text{Hg}^{2+}$  and MeHg to organisms such as bacteria and phytoplankton (Gorski et al., 2008; Gilmour et al., 2011; Luengen et al., 2012).

Mercury accumulation at higher trophic levels can be reduced with the help of increased algal biomass which will help in the dilution of Hg in consumed algal cells. Under increased algal biomass the bloom dilution will reduce the concentration of mercury per cell, as result it would not only reduce dietary input to grazers but also lower bioaccumulation in eutrophic systems that is rich in algal biomass (Pickhardt, Folt, Chen, Klaue, & Blum, 2002). Although some research has suggested that the structure of DOM may have an indirect impact on the rate of Hg photoreactions; however the exact mechanism of DOM is unknown that facilitates mercury reduction (O'Driscoll et al., 2006). A study conducted by Wedyan, Ababneh, and Al-Rousan (2012) established the fact that there is strong affinity between organic matter and mercury which is primarily related to the physical characteristic of these compounds and not to their chemical reactivity, these physical characteristics play the most significant role in the Hg distribution in the sediments. DOM facilitates the mobility of Hg across the column of water and sediment and also results in enhanced sequestration and concentrations of mercury in sediments (Wallschlager et al., 1996; Cossa & Gobeil, 2000). The source of organic matter and dissolved mercury in sediments is the activity of microorganisms. The DOM in these sediments are consumed by different types of bacteria, as the mercury deposited under new layers of sediments and these bacteria live few millimeters under the water interface, a level where oxygen cannot reach. Chemical reactions that involve iron and sulfur are the source of energy for bacteria and these reactions also generate toxic byproduct which is methylmercury (Wedyan, Ababneh, & Al-Rousan, 2012). Accumulation of mercury in fish is a public health concern globally, reason being fishes are the primary source of toxic MeHg to humans. In the next section we will focus and review the existing literature on the effects of dietary exposure to methylmercury amongst the seafood consumer in UAE.

### 3.2 Mercury and Methylmercury Associated Risks

Mercury (Hg) is a pollutant that can be found even in distant areas that are far from the actual emission sources. The level of toxicity is differential and is depended on the form of this metal present in atmosphere and that moves between water, sediment, soil and organisms. Mercury (Hg) can enter the aquatic environment through various pathways. Usually this element enters the aquatic ecosystems through the material discharged and transported in the atmosphere or directly from the watershed (Driscoll et al., 2007). Once mercury enters in the water, it is converted into methylmercury (MeHg) by bacteria this process is known as organic complexing. Methylmercury is a neuro-toxic substance and can cause damages to nerve tissue (US EPA, 2001; Environmental and Health Concerns, 2002).

Methylmercury (MeHg) and inorganic mercury Hg (II) from wet and dry atmospheric deposition enters the aquatic environment directly where they are transported to the shores that is bound to either dissolved organic carbon or to suspended humus or soil, MeHg and Hg (II) can enter the water body by leaching onto the upper soil layers from groundwater flow (Morel et al., 1998; USEPA, 1997; Fitzgerald et al., 2007). In the aquatic system the speciation of mercury is determined by the sequence of complicated biogeochemical processes which includes transportation and transformation (Morel et al., 1998; Fitzgerald et al., 2007). The bioaccumulation of Hg in living organisms especially in fishes poses an environmental mercury threat on the entire food chain (Zhou & Wong, 2000; Kennedy, 2003; Nevado et al., 2003; Hylander & Goodsite, 2006). Where smaller fishes eat the MeHg contaminated microorganisms such as plankton, larger fishes feed on the smaller ones, and so the process goes and the fishes that live long feeding on other fishes accumulate higher levels of MeHg in their tissues (Environmental Protection Agency, 2009).

The level of MeHg presented in water body is influenced by the amount of algae presented in it (Graham, 2002). Studies have shown that in an aquatic environment the concentration of elemental mercury is reduced as elemental mercury is converted to methyl mercury by algae/aquatic plants (Graham, 2002; Green, 2002). When MeHg is transferred from algae to herbivore fish it starts to affect fishes by getting interlocked with the proteins within the tissues (Wooltron, 2002). It is then transferred through the aquatic food chain to and eventually to humans through consumption (Meyers, 1998; Wooltron, 2002).

Studies have shown that prolonged exposure to MeHg has negative effects on the developing brain and the

central nervous system (Castoldi et al., 2008; Newland, Paletz, & Reed, 2008). However, mercury is present in almost every food and it is difficult for humans to survive only on food that do not contain mercury. The usual mercury intake per week from food is 0.3–1.5  $\mu\text{g}/\text{kg}$  body weight (WHO/FAO Provisional tolerable weekly intake (PTWI)). Around 90% of dietary exposure to mercury is through seafood and fish products and 95–100% of MeHg is absorbed from the intestinal tract, perhaps mostly these products contain methylmercury (Patrick, 2002). MeHg is the most toxic form of mercury and is not properly excreted from human body as it is highly lipophilic and therefore considered to be the highest risks to public health (Davidson, Krometis, Al-Harathi, & Gibson, 2012).

### 3.3 Formation of Methylmercury and Its Dietary Exposure in UAE

In UAE around 90 percent of citizens eat fish at least one meal a week, primary Emirati staple diet has historically been the domestically harvested seafood. There are numerous health benefits that have been documented related to the consumption of fish which includes risk reduction of chronic heart disease; however, fish/seafood also act as agents for heavy metals such as mercury, pathogenic microorganisms such as *Vibrio* spp. and other toxins like dioxin estimates of the illness resulting from seafood consumption focus on exposure to mercury. Although numerous metals can result in adverse health effects if consumed in seafood, mercury is generally regarded as of greatest concern. Toxicity to aquatic organisms is majorly caused by the heavy metal contamination of the aquatic environment (Buggiani & Vannucchi, 1980). Al-Ghais (1995) investigated the level of metal (such as copper, cobalt, manganese, lead, nickel) concentration in the heart and liver and mercury concentration in muscle tissue using a large sample size of *Sparus Sarba* a popular fish of UAE that was collected over the period of 4 months and found a heavy mercury concentration in the tissues of these fishes.

#### 3.3.1 Methylmercury Contamination

There have been increasingly serious threats in UAE to aquatic ecosystems and human beings who rely on marine resources for various purposes such as industry, food and also recreation due to heavy metal contamination of marine and coastal environments that are introduced via variety of activities and sources including industrial effluents, sewage and brine discharges, oil pollution and coastal modifications (Naser, 2013). The region that is majorly impacted anthropogenically is the Arabian Gulf amongst other regions in the world. Along the Arabian Gulf 83,600  $\text{km}^2$  is occupied by the UAE, with an estimated population of around 7.5 million (as of 2011). The UAE has a diversified modern economy and resultantly faces similar issues that other industrial nations do in relation to public health and environmental problems due to various industrial activities (Gibson et al., 2013a).

Characteristics that contribute to fragile marine ecosystem of Gulf are shallow, warm water, high evaporation rate, relatively slow surface current, and poor flushing, these may in turn aggravate consequences related to heavy metal contamination (De More, Fowler, Wyse, & Azemard, 2004). A study by Gibson et al. (2013b) shows the level of seafood contamination in fish harvested in Sharjah that was estimated using mean and confidence interval values showing high mercury contamination. As per the findings MeHg node was found in continuous triangular distribution (minimum-0.033, mode-0.068 and maximum-0.098) presented in units of  $\text{mg}/\text{kg}$ , the results were particularly observed in *Redspot emperor*.

#### 3.3.2 Dietary Exposure

According to Gibson et al. (2013b), UAE resident could be at the risk of over dietary exposure to methylmercury from consumption of seafood as their level of mercury dietary exposure is much higher than the reference dose due to consumption of fish as maintained by the U.S. Environmental Protection Agency.

Table 1. Estimated accountability of disease due to exposure to methylmercury in the UAE (Gibson &amp; Farah, 2012)

Exposure route	Risks evaluated	Exposure indicators	Adverse health conditions	Attributable fatalities in 2008 (95% CI)	Attributable health care facility visits in 2008 (95% CI)
Food (eating)	Seafood contamination	Methylmercury	Neurological disorders	4(0, 10)	27000(0, 67,000)

Study conducted by Davidson, Krometis, Al-Harhi, and Gibson (2012) using stochastic model estimated that 1 in 5 run into daily risk of dietary exposure to methylmercury through consumption of seafood which is much higher and exceeds the Food and Agriculture Organization and World Health Organization (FAO/WHO) provisional tolerable intake on a weekly basis. The benchmark for organic MeHg exposure, that they used was WHO/FAO Provisional tolerable weekly intake (PTWI) of 1.6 µg/kg body weight (0.00023 mg/(kg day)) (WHO Technical Report, 2003). Table 2 shows the parameters of varying effects of seafood sensitivity.

Table 2. Sensitivity effects of varying seafood parameters

Parameter	Base Case	MeHg Contamination		Serving Size		Daily Consumption Frequency	
		-10%	10%	-10%	10%	-10%	10%
Average probability* of exceeding MeHg benchmark**	1:05	1:05	1:04	1:05	1:04	N/A	N/A
Daily cases exceeding MeHg benchmark	451,019	375,140	514,800	375,439	514,681	401,273	490,445

\*For instance, 1:05 suggests 1 in 5 consumers of seafood in the UAE will exceed the benchmark of MeHg on any given day. \*\*FAO/WHO provisional tolerable weekly intake (PTWI) benchmark of 1.6 µg/kg body weight as daily max: 0.00023 mg/kg body weight (WHO Technical Report, 2003). Adopted from (Davidson, Krometis, Al-Harhi, & Gibson, 2012).

Potential human illness symptoms due to lead, mercury or manganese contamination and dietary exposure include neurological effects (including loss of IQ amongst children whose mothers are exposed to MeHg); exposure to other metals such as cadmium can lead to cancers; and cardiovascular disease results from exposure to arsenic (U.S. Environmental Protection Agency, 2008).

As reported, these high probabilities of exceeding levels of reference dose of PTWI of MeHg through dietary exposure of seafood consumers in UAE may give way to additional regulations and policies to reduce this metal exposure. A level of mercury that can be consumed without negative effects by humans over a lifetime is a reference dose (Baker, 2003). Though some researchers still have not drawn to close a possible solution related to MeHg contamination and they continue to attempt to stipulate a standard for safe exposure to mercury levels so that useful advisories on consumption of fish and seafood could be issued effectively (Myers, 1998; Graham, 2003). Nonetheless, any measures taken to limit the exposure of MeHg must be adequately balanced with the various considerable health benefits that are related to the consumption of seafood, particularly keeping the rising levels of obesity in mind that has been recently observed amongst the population of UAE (UAE-UNC, 2010).

#### 4. Discussion

This review reveals that mercury contamination of marine environment in UAE can pose considerable threat to public health. The dominant and major source of mercury exposure is considered to be dietary intake amongst the general population. The inorganic Hg is converted into methyl mercury (organic Hg) once it is released in the aquatic environment where it primarily accumulates in shellfish and fish at approximate concentrations of 1000–10,000 times higher than its presence in other form of foods (Rice et al., 2000). The risk of dietary exposure to Hg through the consumption of shellfish and is subjected to the level of mercury in the seafood and the amount of fish and shellfish eaten on daily/weekly basis. Hence, the Environmental Protection Agency (EPA)

and food and drug administration (FDA) have specially advised to avoid consuming fish with high mercury levels to young children, nursing mothers, pregnant women and women who are likely to get pregnant (Mahaffey, 2004; USFDA, 2006). The major concern is the exceeding WHO PTWI levels of dietary exposure to MeHg in UAE and its neurotoxicity effect that can hamper the development of brain and lead to neurological disorders.

Though comprehensive studies have been conducted to determine the bio-accumulation and toxicity of these metals in marine environment and especially in fishes in America and Europe (Pedan et al., 1973; Nickless et al., 1972; Hardisty et al., 1974; Badsha & Sainsbury, 1978), very few, rather limited studies are available in UAE that determine the level of metal contamination in the marine flora and fauna. There is a need for more Empirical studies in UAE that need to be conducted in future to evaluate specific health risks associated with estimated MeHg exposure to better understand the relationship of dose and its related response in terms of public health outcomes (Steenland, 1996; Cohen et al., 2005). There are only limited studies available in regards to the concentrations of these metal contaminants that exist in coastal waters of UAE, and fewer corresponding data on methylmercury concentrations in seafood. Due to this limitation of UAE-specific information, it is challenging to assess the exact nature of public health risks specifically posed by MeHg dietary exposure through seafood (Al Zarooni & Elshorbagy, 2006). Thus, this review may be helpful in providing information for future investigation on the alarming methylmercury contamination of aquatic ecosystem in UAE.

## References

- Aiken, G., & Costaris, E. (1995). Soil and hydrology: their effect on NOM. *J. A WWA*, 87, 36-45.
- Al Zarooni, M., & Elshorbagy, W. (2006). Characterization and assessment of Al Ruwais refinery wastewater. *J. Hazard.Mater.*, 136, 398-405. <http://dx.doi.org/10.1016/j.jhazmat.2005.09.060>
- Al-Ghais, S. M. (1995). Heavy metal concentrations in the tissue of Sparus sarba Forskål, 1775 from the United Arab Emirates. *Bull Environ Contam Toxicol*, 55(4), 581-7. <http://dx.doi.org/10.1007/BF00196039>
- Andren, A. W., & Harriss, R. C. (1975). Observations on the association between mercury and organic matter dissolved in natural waters. *Gcochimica Cosmochimiea Acts*, 39, 1253-1257.
- Badsha, K. S., & Sainsbury, M. (1978). Some aspects of the biology and heavy metal accumulation of the fish *Liparis liparis* in the Severn Estuary. *Estuarine Coastal Mar. Sci.*, 7, 381-391. [http://dx.doi.org/10.1016/0302-3524\(78\)90090-7](http://dx.doi.org/10.1016/0302-3524(78)90090-7)
- Baeyens, W., & Leermakers, M. (1996). Particulate, dissolved and methylmercury budgets for the Scheldt estuary (Belgium and the Netherlands). In W. Baeyens, R. Ebinghaus, & O. Vasiliev (Eds.), *Global and Regional Mercury Cycles: Sources, fluxes and mass balances* (pp. 285-301). Kluwer Academic Publishers.
- Baker, B. (2003). *Mercury in Fish: Agencies Work to Find Common Ground*. Retrieved from <http://web9.epnet.com/Deliveryprintsave.asp>.
- Bernhardt, E. S., & Likens, G. E. (2002). Dissolved organic carbon enrichment alters nitrogen dynamics in a forest stream. *Ecology*, 83, 1689-1700. [http://dx.doi.org/10.1890/0012-9658\(2002\)083\[1689:DOCEAN\]2.0.CO;2](http://dx.doi.org/10.1890/0012-9658(2002)083[1689:DOCEAN]2.0.CO;2)
- Bortoli, A. M., Gerotto, M., Marchiori, M., Palonta, M., & Troncon, A. (1995). Analytical problems in mercury analysis of seafood. *Annali dell'Istituto superiore di sanita*, 31(3), 359-62.
- Bothwell, M. L., Sherbot, D., Roberge, A. C., & Daley, R. J. (1993). Influence of natural ultraviolet radiation on lotic periphytic diatom community growth, biomass accrual, and species composition: short-term versus long-term effects. *Journal of Phycology*, 29, 24-35. <http://dx.doi.org/10.1111/j.1529-8817.1993.tb00276.x>
- Buffle, J., Chalmers, R. A., Masson, M. R., & Midgley, D. (1988). *Complexation Reactions in Aquatic Systems: An Analytical Approach* (1st ed.). Chichester: Halsted Press.
- Buggiani, S. S., & Vannucchi, C. (1980). Mercury and lead concentrations in some species of fish from Tuscan Coast (Italy). *Bull. Environ. Contam. Toxicol.*, 25, 90-92. <http://dx.doi.org/10.1007/BF01985493>
- Cassen, R. H. (1978). Current trends in population change and their causes. *Popul Dev Rev*, 4(2), 331-353. <http://dx.doi.org/10.2307/1972285>
- Castoldi, A. F., Johansson, C., Onishchenko, N., Coccini, T., Roda, E., Vahter, M., . . . Manzo, L. (2008). Human developmental neurotoxicity of methyl mercury: Impact of variables and risk modifiers. *Regul Toxicol Pharm*, 51(2), 201-214. <http://dx.doi.org/10.1016/j.yrtph.2008.01.016>
- Cauwet, G. (1978). Organic chemistry of seawater particulates concepts and developments. *Ocean01 Acta*, 1(1), 99-105.

- Cohen, J. T., Bellinger, D. C., Connor, W. E., Kris-Etherton, P. M., Lawrence, R. S., Savitz, D.A., . . . Gray, G. M. (2005). A quantitative risk-benefit analysis of changes in population fish consumption. *Am. J. Prev. Med.*, *29*, 325-334.
- Cossa, D., & Gobeil, C. (2000). Mercury speciation in the lower St. Lawrence Estuary. *Can. J. Fish. Aquat. Sci.*, *57*, 138-147. <http://dx.doi.org/10.1139/f99-237>
- Davidson, C. A., Krometis, L. A., Al-Harhi, S. S., & Gibson, J. M. (2012). Foodborne exposure to pesticides and methyl mercury in the United Arab Emirates. *Risk Anal.*, *2*(3), 381-94. <http://dx.doi.org/10.1111/j.1539-6924.2011.01679.x>
- De Mora, S., Fowler, S. W., Wyse, E., & Azemard, S. (2004). Distribution of heavy metals in marine bivalves, fish and coastal sediments in the Gulf and Gulf of Oman. *Marine Pollution Bulletin*, *49*, 410-424. <http://dx.doi.org/10.1016/j.marpolbul.2004.02.029>
- Dong, W., Liang, L. Brooks, S., Southworth, G., & Gu, B. (2010). Roles of dissolved organic matter in the speciation of mercury and methyl mercury in a contaminated ecosystem in Oak Ridge, Tennessee. *Environmental Chemistry*, *7*, 94-102. <http://dx.doi.org/10.1071/EN09091>
- Drexel, R. T., Haitzer, M., Ryan, J. N., Aiken, G. R., & Nagy, K. L. (2002). Mercury (II) sorption to two Florida Everglades peats: evidence for strong and weak binding and competition by dissolved organic matter released from the peat. *Environ. Sci. Technol.*, *36*, 4058-4064. <http://dx.doi.org/10.1021/es0114005>
- Driscoll, C. T., Blette, V., Yan, C., Schofield, C. L., Munson, R., & Holsapple, J. (1995). The role of dissolved organic carbon in the chemistry and bioavailability of mercury in remote Adirondack lakes. *Water, Air, and Soil Pollution*, *80*(1-4), 499-508.
- Driscoll, C. T., Han, Y. J., Chen, C. Y., Evers, D. C., Lambert, K. F., Holsen, T. M., . . . Munson, R. K. (2007). Mercury contamination in forest and freshwater ecosystems in the northeastern United States. *BioScience*, *57*, 17-28. <http://dx.doi.org/10.1641/B570106>
- Environmental and Health Concerns. (2002). *Environment Canada*. Retrieved from <http://www.ec.gc.ca/mercury/ehc-hs-e.html>
- Environmental Protection Agency. (2009). *National Listings Fish Advisories*. Retrieved from <http://www.epa.gov/waterscience/fish/advisories/>
- Fellman, J. B., Spencer, R. G. M., Hernes, P. J., Edwards, R. T., D'Amore, D. V., & Hood, E. (2010). The impact of glacier runoff on the biodegradability and biochemical composition of terrigenous dissolved organic matter in near-shore marine ecosystems. *Marine Chemistry*, *121*(1-4), 112-122. <http://dx.doi.org/10.1016/j.marchem.2010.03.009>
- Fitzgerald, W. B., & Lyon, S. (1973). Organic mercury compounds in coastal waters. *Nature*, *242*, 452-453. <http://dx.doi.org/10.1038/242452a0>
- Fitzgerald, W. F., Lamborg, C. H., & Hammerschmidt, C. R. (2007a). Marine biogeochemical cycling of mercury. *Chem. Rev.*, *107*, 641-662. <http://dx.doi.org/10.1021/cr050353m>
- Fitzgerald, W. F., Lamborg, C. H., Heinrich, D. H., & Karl, K. T. (2007b). *Geochemistry of mercury in the environment, treatise on geochemistry*. Pergamon: Oxford.
- Fleming, L. E., Broad, K., Clement, A., Dewailly, E., Elmir, S., Knap, A., . . . Walsh, P. (2006). Oceans And Human Health: Emerging Public Health Risks In The Marine Environment. *Marine Pollution Bulletin*, *53*(10-12), 545-560. <http://dx.doi.org/10.1016/j.marpolbul.2006.08.012>
- Gerbig, C. A., Ryan, J. N., & Aiken, G. R. (2011). The Effects of Dissolved Organic Matter on Mercury Biogeochemistry. *Environmental Chemistry and Toxicology of Mercury*, 259-292. <http://dx.doi.org/10.1002/9781118146644.ch8>
- Gibson, J. M., & Farah, Z. S. (2012). Environmental risks to public health in the United Arab Emirates: a quantitative assessment and strategic plan. *Environmental Health Perspectives*, *120*(5), 681-686. <http://dx.doi.org/10.1289/ehp.1104064>
- Gibson, J. M., Brammer, A. S., Davidson, C. A., Folley, T., Launay, F. J. P., & Thomsen, J. T. W. (2013b). Burden of Disease from Produce and Seafood Contamination. *Environmental Science and Technology Library*, *24*, 307-348. [http://dx.doi.org/10.1007/978-94-007-5925-1\\_11](http://dx.doi.org/10.1007/978-94-007-5925-1_11)
- Gibson, J. M., Thomsen, J., Launay, F., Harder, E., & DeFelice, N. (2013a). Deaths and Medical Visits Attributable



- to Environmental Pollution in the United Arab Emirates. *PLoS ONE*, 8(3), e57536. <http://dx.doi.org/10.1371/journal.pone.0057536>
- Gilmour, C. C., Elias, D. A., Kucken, A. M., Brown, S. D., Palumbo, A. V., Schadt, C. W., & Wall, J. D. (2011). Sulfate-reducing bacterium *Desulfovibrio desulfuricans* ND132 as a model for understanding bacterial mercury methylation. *Appl. Environ. Microbiol.*, 77, 3938-3951. <http://dx.doi.org/10.1128/AEM.02993-10>
- Gimbel, B. (2007). The richest city in the world. *Fortune*, 155(5), 168.
- Gorski, P. R., Armstrong, D. E., Hurley, J. P., & Krabbenhoft, D. P. (2008). Influence of natural dissolved organic carbon on the bioavailability of mercury to a freshwater alga. *Environ. Pollut.*, 154, 116-123. <http://dx.doi.org/10.1016/j.envpol.2007.12.004>
- Graham, S. (2003). *Organisms in Algae-Rich Lakes May Absorb Less Mercury*. Scientific American. Retrieved from <http://www.scientificamerican.com/article.cfm?id=organisms-in-algae-rich-l>
- Green, K. (2003). *Murky Ponds Minimize Mercury*. Retrieved from [http://bric.postech.ac.kr/science/9/now/02\\_3now/2032lb.html](http://bric.postech.ac.kr/science/9/now/02_3now/2032lb.html)
- Haitzer, M., Aiken, G. R., & Ryan, J. N. (2002). Binding of Mercury(II) to dissolved organic matter: the role of mercury to DOM concentration ratio. *Environ. Sci. Technol.*, 36, 3564-3570.
- Haitzer, M., Aiken, G. R., & Ryan, J. N. (2003). Binding of Mercury (II) to Aquatic Humic Substances: Influence of pH and Source of humic substances. *Environ. Sci. Technol.*, 37, 2436-2441. <http://dx.doi.org/10.1021/es026291o>
- Hansell, D. A. (2002). DOC in the global ocean carbon cycle. In D. A. Hansell, & C. A. Carlson (eds.), *Biogeochemistry of marine dissolved organic matter* (pp. 509-546). Academic Press.
- Hardisty, M. W., Huggins, R. J., Harter, S., & Sansbury, M. (1974). Ecological implications of heavy metals in fish from the Sever estuary. *Marine Pollution Bulletin*, 5, 12-15. [http://dx.doi.org/10.1016/0025-326X\(74\)90027-7](http://dx.doi.org/10.1016/0025-326X(74)90027-7)
- Hill, W. R., Dimick, S. M., McNamara, A. E., & Branson, C. A. (1997). No effects of ambient UV radiation detected in periphyton and grazers. *Limnology and Oceanography*, 42, 769-774. <http://dx.doi.org/10.4319/lo.1997.42.4.0769>
- Hurley, J. P., Benoit, J. M., Babiarz, C. L., Shafer, M. M., Andren, A. W., Sullivan, J. R., . . . Webb, D. A. (1995). Influences of watershed characteristics on mercury levels in Wisconsin rivers. *Environ. Sci. Technol.*, 29, 1867-1875. <http://dx.doi.org/10.1021/es00007a026>
- Hylander, L. D., & Goodsite, M. E. (2006). Environmental costs of mercury pollution. *Science of the Total Environment*, 368(1), 352-370. <http://dx.doi.org/10.1016/j.scitotenv.2005.11.029>
- Kelly, D. J., Clare, J. J., & Bothwell, M. L. (2001). Attenuation of solar ultraviolet radiation by dissolved organic matter alters benthic colonization patterns in streams. *Journal of the North American Benthological Society*, 20, 96-108. <http://dx.doi.org/10.2307/1468191>
- Kennedy, C. J. (2003). Uptake and accumulation of mercury from dental amalgam in the common goldfish, *Carassius auratus*. *Environmental Pollution*, 121(3), 321-326. [http://dx.doi.org/10.1016/S0269-7491\(02\)00271-3](http://dx.doi.org/10.1016/S0269-7491(02)00271-3)
- Kolka, R. K., Grigal, D. F., Verry, E. S., & Nater, E. A. (1999). Mercury and organic carbon relationships in streams draining forested upland/peatland watersheds. *J. Environ. Qual.*, 28, 766-775. <http://dx.doi.org/10.2134/jeq1999.00472425002800030006x>
- Krabbenhoft, D. (2003). *Mercury Contamination of Aquatic Ecosystems—Summary of findings on Sources, Cycling, and Trends*. Retrieved from <http://www.e3ventures.com/mercury.PDF.krabbenhoft1.pdf>
- Lamborg, C. H., Fitzgerald, W. F., Skoog, A., & Visscher, P. T. (2004). The abundance and source of mercury-binding organic ligands in Long Island Sound. *Mar. Chem.*, 90(1-4), 51-163. <http://dx.doi.org/10.1016/j.marchem.2004.03.014>
- Lawrence, A. L., & Mason, R. P. (2001). Factors controlling the bioaccumulation of mercury and methylmercury by the estuarine amphipod *Leptocheirus plumulosus*. *Environmental Pollution*, 11, 217-231. [http://dx.doi.org/10.1016/S0269-7491\(00\)00072-5](http://dx.doi.org/10.1016/S0269-7491(00)00072-5)
- Lindberg, S., & Harriss, R. (1974). Mercury-organic matter associations in estuarine sediments and interstitial water. *Environmental Science & Technology*, 8, 459. <http://dx.doi.org/10.1021/es60090a009>

- Loux, N. T. (1998). An assessment of mercury-species dependent binding with natural organic carbon. *Chem. Spec. Bioavail*, 10, 127-136. <http://dx.doi.org/10.3184/095422998782775754>
- Luengen, A. C., Fisher, N. S., & Bergamaschi, B. A. (2012). Dissolved organic matter reduces algal accumulation of methylmercury. *Environ. Toxicol. Chem.*, 31(8), 1712-1719. <http://dx.doi.org/10.1002/etc.1885>
- Mahaffey, K. (2004). *Methylmercury: Epidemiology Update. Fish Forum, San Diego*. Retrieved from [www.epa.gov/waterscience/fish/forum/2004/presentations/monday/mahaffey.pdf](http://www.epa.gov/waterscience/fish/forum/2004/presentations/monday/mahaffey.pdf)
- Mason, R. P., Reinfelder, J. R., & Morel, F. M. M. (1995). *Bioaccumulation of mercury and methylmercury. Water, Air, and Soil Pollution*, 80, 915-921. <http://dx.doi.org/10.1007/BF01189744>
- Mason, R. P., Reinfelder, J. R., & Morel, F. M. M. (1996). The uptake, toxicity, and trophic transfer of mercury in a coastal diatom. *Environ. Sci. Technol.*, 30, 1835-1845. <http://dx.doi.org/10.1021/es950373d>
- McKnight, D. M., & Aiken, R. (1998). Sources and age of aquatic humic. In D. O. Hessen, & L. J. Tranvik (Eds.), *Aquatic Humic Substances: Ecology and Biogeochemistry* (pp. 9-37). Germany: Springer.
- Meili, M., Iverfeldt, A., & Hakanson, L. (1991). *Mercury in the surface water of Swedish forest lakes—Concentrations, speciation, and controlling factors. Water, Air, and Soil Pollution*, 56, 439-453. <http://dx.doi.org/10.1007/BF00342290>
- Mierle, G., & Ingram, R. (1991). The role of humic substances in the mobilization of mercury from watersheds. *Water Air & Soil Pollution*, 56, 349-357. <http://dx.doi.org/10.1007/BF00342282>
- Mopper, K., & Degens, E. T. (1979). Organic carbon in the ocean: nature and cycling. In B. Bolin, E. T. Degens, S. Kempe, & P. Ketner (Eds.), *The global carbon cycle* (pp. 293-315). New York: John Wiley & Sons.
- Morel, F. M. M., Kraepiel, A. M. L., & Amyot, M. (1998). *The chemical cycle and bioaccumulation of mercury. Annual Review of Ecology and Systematics*, 29, 543-566. <http://dx.doi.org/10.1146/annurev.ecolsys.29.1.543>
- Moye, H. A., Miles, C. J., Philips, E. J., Sargent, B., & Merritt, K. K. (2002). Kinetics and uptake mechanisms for monomethylmercury between freshwater algae and water. *Environ. Sci. Technol.*, 36, 3550-3555. <http://dx.doi.org/10.1021/es011421z>
- Myers, G. (1998). *Studying the Effects of Methyl Mercury*. Retrieved from [http://www.senate.gov/~epw/105th/mye\\_10-1.htm](http://www.senate.gov/~epw/105th/mye_10-1.htm)
- Myers, G. (2003). *Statement by the University of Rochester Research Team Studying the Effects of MeHg*. Read Before the Senate Subcommittee on Clean Air, Wetlands, Private Property and Nuclear Safety. Committee on Environment and Public Works. Retrieved from [http://www.senate.gov/~epw/105th/mye\\_10-1.htm](http://www.senate.gov/~epw/105th/mye_10-1.htm)
- Naser, H. A. (2013). Assessment and management of heavy metal pollution in the marine environment of the Arabian Gulf: a review. *Mar. Pollut. Bull.*, 72(1), 6-13. <http://dx.doi.org/10.1016/j.marpolbul.2013.04.030>
- Nevado, J. J. B., Bermejo, L. F. G., & Martn-Doimeadios, R. C. R. (2003). Distribution of mercury in the aquatic environment at Almaden, Spain. *Environmental Pollution*, 122(2), 261-271. [http://dx.doi.org/10.1016/S0269-7491\(02\)00290-7](http://dx.doi.org/10.1016/S0269-7491(02)00290-7)
- Newland, M. C., Paletz, E. M., & Reed, M. N. (2008). Methylmercury and nutrition: Adult effects of fetal exposure in experimental models. *Neurotoxicology*, 29(7), 783-801. <http://dx.doi.org/10.1016/j.neuro.2008.06.007>
- Nickless, G., Stenner, R., & Terrille, N. (1972). Distribution of cadmium, lead and zinc in Bristol Channel. *Mar. Pollut. Bull.*, 3(12), 188-190. [http://dx.doi.org/10.1016/0025-326X\(72\)90267-6](http://dx.doi.org/10.1016/0025-326X(72)90267-6)
- Nriagu, J. O. (1994). Mechanistic steps in photoreduction of mercury in natural water. *Sci Total Environ*, 154, 1-8. [http://dx.doi.org/10.1016/0048-9697\(94\)90608-4](http://dx.doi.org/10.1016/0048-9697(94)90608-4)
- O'Driscoll, N. J., Siciliano, S. D., Lean, D. R. S., & Amyot, M. (2006). *Gross photo reduction kinetics of mercury in temperature freshwater lakes and rivers: Application to a general model of DGM dynamics. Environ. Sci. Technol.*, 40, 837-843. <http://dx.doi.org/10.1021/es051062y>
- Patrick, L. (2002). Mercury toxicity and antioxidants: Part 1: role of glutathione and alpha-lipoic acid in the treatment of mercury toxicity. *Altern. Med. Rev.*, 7, 456-471.
- Peden, J. D., Crothers, J. H., Waterfall, C. E., & Beasley, J. (1973). Heavy metals in Somerset marine organisms. *Mar. Pollut. Bull.*, 4, 7-10. [http://dx.doi.org/10.1016/0025-326X\(73\)90022-2](http://dx.doi.org/10.1016/0025-326X(73)90022-2)
- Pickhardt, P. C., & Fisher, N. S. (2007). Accumulation of inorganic and methylmercury by freshwater phytoplankton in two contrasting water bodies. *Environ. Sci. Technol.*, 41, 125-131.

- <http://dx.doi.org/10.1021/es060966w>
- Pickhardt, P. C., Folt, C. L., Chen, C. Y., Klaue, B., & Blum, J. D. (2002). Algal blooms reduce the uptake of toxic methylmercury in freshwater food webs. *Proc. Natl. Acad. Sci. USA*, 99(7), 4419-23. <http://dx.doi.org/10.1073/pnas.072531099>
- Price, A. R. G., Sheppard, C. R. C., & Roberts, C. M. (1993). The Gulf: its biological setting. *Marine Pollution Bulletin*, 27, 5-15. [http://dx.doi.org/10.1016/0025-326X\(93\)90004-4](http://dx.doi.org/10.1016/0025-326X(93)90004-4)
- Ravichandran, M. (2004). Interactions between mercury and dissolved organic matter—a review. *Chemosphere*, 55(3), 319-331. <http://dx.doi.org/10.1016/j.chemosphere.2003.11.011>
- Rice, G., Swartout, J., Mahaffey, K., & Schoeny, R. (2000). Derivation of US EPA's oral Reference Dose (RfD) for methylmercury. *Drug Chem. Toxicol.*, 23(1), 41-54. <http://dx.doi.org/10.1081/DCT-100100101>
- Shanley, J. B., Schuster, P. F., Reddy, M. M., Roth, D. A., Taylor, H. E., & Aiken, G. R. (2002). Mercury on the move during snowmelt in Vermont. *EOS*, 83(5), 45-48. <http://dx.doi.org/10.1029/2002EO000031>
- Sleighter, R. L., Liu, Z., Xue, J., & Hatcher, P. G. (2010). Multivariate statistical approaches for the characterization of dissolved organic matter analyzed by ultrahigh resolution mass spectrometry. *Environ. Sci. Technol.*, 44(19), 7576-82. <http://dx.doi.org/10.1021/es1002204>
- Slowey, A. J. (2010). Rate of formation and dissolution of mercury sulfide nanoparticles: The dual role of natural organic matter. *Geochimica et Cosmochimica Acta*, 74(16), 4693-4708. <http://dx.doi.org/10.1016/j.gca.2010.05.012>
- Stenson, A. C., Marshall, A. G., & Cooper, W. T. (2003). Exact masses and chemical formulas of individual Suwannee River fulvic acids from ultrahigh resolution electrospray ionization Fourier transform ion cyclotron resonance mass spectra. *Anal. Chem.*, 75, 1275-1284.
- Strober, Q. J., Jones, R. D., & Scheidt, D. J. (1995). Ultra-trace level mercury in the place Everglades ecosystem: a multi media canal pilot study. *Water Air & Soil Pollution*, 80, 991-1001. <http://dx.doi.org/10.1007/BF01189753>
- Tuchman, N. C. (1996). The role of heterotrophy in algae. In R. Stevenson., M. L. Bothwell, & R. L. Lowe (Eds.), *Algal Ecology: Freshwater Benthic Ecosystems* (pp. 299-316). San Diego, CA, U.S.A: Academic press.
- U.S. Energy Information Administration. (2010). *United Arab Emirates: Country Analysis Brief*. Washington, DC: U.S. Department of Energy, Energy Information Administration. Retrieved from <http://205.254.135.24/countries/country-data.cfm?fips=TC>
- U.S. Environmental Protection Agency. (2008). *Integrated Risk Information System*. Retrieved from <http://www.epa.gov/iris>
- United Nations. (2007). *World Population Prospects: The 2006 Revision: Highlights*. New York: United Nations, Department of Economic and Social Affairs, Population Division. Retrieved from [http://www.un.org/esa/population/publications/wpp2006/WPP2006\\_Highlights\\_rev.pdf](http://www.un.org/esa/population/publications/wpp2006/WPP2006_Highlights_rev.pdf)
- United States Food and Drug Administration. (2006). *Mercury Levels in Commercial Fish and Shellfish*. Retrieved from <http://www.cfsan.fda.gov/~frf/sea-mehg.html>
- USEPA. (1997). *Mercury Study Report to Congress Volume IV: An Assessment of Exposure to Mercury In The United States*. Retrieved from <http://www.epa.gov/ttn/caaa/t3/reports/volume4.pdf>
- USEPA. (2001). *Method 1630: Methyl Mercury in Water by Distillation, Aqueous Ethylation, Purge and Trap, and CVAFS*. Retrieved from <http://www.brooksrand.com/Documents/1630.pdf>
- Vinebrooke, R. D., & Leavitt, P. R. (1998). *Direct and interactive effects of allochthonous dissolved organic matter, inorganic nutrients, and ultraviolet radiation on an alpine littoral food web*. *Limnology and Oceanography*, 43, 1065-1081. <http://dx.doi.org/10.4319/lo.1998.43.6.1065>
- Wallschlager, D., Desai, M. V. M., & Wilken, R. D. (1996). The role of humic substances in the aqueous mobilization of mercury from contaminated floodplain soils. *Water Air & Soil Pollution*, 90, 507-520. <http://dx.doi.org/10.1007/BF00282665>
- Waples, J. S., Nagy, K. L., Aiken, G. R., & Ryan, J. N. (2005). Dissolution of cinnabar (HgS) in the presence of natural organic matter. *Geochimica et Cosmochimica Acta*, 69(6), 1575-1588. <http://dx.doi.org/10.1016/j.gca.2004.09.029>
- Watras, C. J., Morrison, K. A., & Host, J. S. (1995). Concentration of mercury species in relationship to other

- site-specific factors in the surface waters of northern Wisconsin lakes. *Limnol. Oceanogr.*, 40, 553-565. <http://dx.doi.org/10.4319/lo.1995.40.3.0556>
- Wedyan, M. A., Ababneh, F. A., & Al-Rousan, S. (2012). The correlations between mercury speciation and dissolved organic matter in the sediment of the red sea. *Am. J. Environ. Sci.*, 8, 403-411. <http://dx.doi.org/10.3844/ajessp.2012.403.411>
- WHO Technical Report Series. (2003). *Evaluation of certain food additives and contaminants: Sixtyfirst report of the joint FAO/WHO Expert Committee on Food Additives*. Retrieved from [http://whqlibdoc.who.int/trs/WHO\\_TRS\\_922.pdf](http://whqlibdoc.who.int/trs/WHO_TRS_922.pdf)
- WHO. (2007). *Country cooperation strategy at a glance: UAE*. Retrieved from [http://www.who.int/countryfocus/cooperation\\_strategy/ccsbrief\\_are\\_en.pdf](http://www.who.int/countryfocus/cooperation_strategy/ccsbrief_are_en.pdf)
- Wooltron, E. (2002). *Facts on Mercury and Fish Consumption*. *Canadian Medical Association journal*, 167(8), 897.
- Wright, A. L., & Reddy, K. R. (2009). *Dissolved Organic Matter in Wetlands*. Retrieved from <http://edis.ifas.ufl.edu/ss507>
- Xenopoulos, M. A., & Schindler, D. W. (2001). Physical factors determining ultraviolet flux into ecosystems. In C. S. Cockell, & A. R. Blaustein (Eds.), *Ecosystems, Evolution and UV Radiation* (pp. 36-62). New York: Springer.
- Zhou, H. Y., & Wong, M. H. (2000). Mercury accumulation in freshwater fish with emphasis on the dietary influence. *Water Research*, 34(17), 4234-4242. [http://dx.doi.org/10.1016/S0043-1354\(00\)00176-7](http://dx.doi.org/10.1016/S0043-1354(00)00176-7)

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