Elephants or Excrement? Comparison of the Power of Two Survey Methods for Elephants in West African Savanna

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Abstract

Statistical power is the key issue when evaluating wildlife trends. Wildlife managers in West Africa lack light aeroplanes and must often estimate wildlife trends on foot, either by counting the animals themselves or by counting their dung. We compared direct ground counts of elephants and dung counts in two consecutive years in Nazinga Game Ranch in southern Burkina Faso, an area of savanna woodland and shrub savanna. Our purpose was to determine which method was most likely to detect a trend in elephant numbers. The direct ground counts each covered 680 km of transect but sightings were too few to give good estimates of elephant numbers. The dung counts each covered 54 km of transect recorded very large numbers of dung-piles and returned estimates with smaller coefficients of variation. Consequently dung counts had greater statistical power for detecting trends. Although dung counts have been mostly applied in forests, they should also be used in open habitats where information on long-term trends is sought.

Keywords: elephant, survey, dung, Burkina Faso, faecal count

1. Introduction

Management of wildlife populations requires estimates of population size and trends. The key issue in detecting trends is statistical power: the probability that a change in numbers will be detected, or, more formally, the probability that the null hypothesis will be rejected when it is false (Taylor & Gerrodette, 1993). Maximum power can often be achieved by increasing sample size, but that is expensive and most wildlife managers operate under severe budgetary constraints. The challenge for the wildlife manager is to find a method that, when used regularly, balances cost with power. Here we compare two methods in the Nazinga Game Ranch, in the West African savanna, for monitoring elephant trends.

Large mammal populations in West Africa are often at low density. Frequently animals are scattered or live in groups that vary in size. They are difficult to see in the field because of their coat color and behavior (Jachmann, 2001). Furthermore, visibility is poor in thick vegetation such as savanna woodland. With direct ground counts, one often sees no animals in many transects, while a few transects record many sightings. These features of herbivore biology and dispersion mean that direct ground counts produce estimates with large variances and wide confidence limits (Jachmann, 1991; Barnes, 2002). The same is true of many aerial sample counts. When counting sparse elephant populations (<0.5 km⁻²), aerial sample counts have poor detection efficiency and return inaccurate estimates with low precision (Bouché, Lejeune, & Vermeulen, 2012). Such estimates lack the statistical power to detect changes in numbers (Taylor & Gerrodette, 1993).

Light aeroplanes are much less common in West Africa than elsewhere on the continent. In addition, aviation fuel is not readily available in most protected areas and flying is often restricted because of security considerations. Consequently aerial surveys, whether total or sample counts, are more difficult and expensive to arrange than in eastern or southern Africa. Therefore wildlife managers in West Africa are often obliged to count animals while walking along transects ("direct ground counts"). A case in point is the Nazinga Game Ranch, in the savanna zone of southern Burkina Faso. Over the years numerous counts of elephants, using different methods, have been conducted. But the large variations between counts, even between counts with similar

methods, mean that managers have no clear idea of the trend in elephant numbers. There is a need for a counting method that is cheap, easy for the park staff to conduct once a year and that will provide consistent estimates to reveal trends in the elephant population.

Large herbivores produce large quantities of dung. Dung-piles, whether of elephants, *Loxodonta africana*, or antelopes, are more abundant and most likely to be detected and counted than the animals that produced them. Yet, until now, with a few notable exceptions (Jachmann, 1991, 2001; Jachmann & Bell, 1979, 1984), many people still have the impression that they should only be used when the animals cannot be seen, for example in dense forests. This despite the fact that the reliability of dropping counts to estimate animal numbers has been well documented (Neff, 1968; Jachmann & Bell, 1984; Johnson & Jarman, 1987; Jachmann, 1991; Barnes, 2001; Karels, Koppel, & Hik, 2004). A review of published accounts of dung counts of vertebrates raging in size from lizards to elephants showed that they gave estimates of animal numbers similar to those from other methods (Barnes, 2001). Dung counts give estimates that are as accurate and precise, or more precise, that aerial sample surveys (Barnes, 2001, 2002).

In this paper we compare direct ground counts and dung counts in order to determine the method with the greater statistical power for detecting trends in elephant numbers in a West African savanna.

2. Study Site

The Nazinga Game Ranch lies in southern Burkina Faso and covers about 975 km² (Figure 1). The area is relatively flat with a mean altitude of 300 m (Spinage, 1984). It is traversed by the valley of the Sissili River and two of its tributaries, all of which flow seasonally. There is a single dry season from October to May and a rainy season from June to September. The mean annual rainfall is about 900 mm. The vegetation is characterized by tall grasses, trees and shrubs with the main vegetation types being riverine forest, savanna woodlands and shrub savanna (Guinko, 1985).

3. Methods

3.1 Elephant Counts 1982–2008

Since 1982 numerous estimates of elephants have been made. However different methods have been used: aerial sample count (one count only), aerial total counts, direct counts on the ground and dropping counts. Since 1985 an annual direct ground count of all large mammals has been made in the dry season using the line-transect method.

3.2 Direct Ground Counts in 2007 and 2008

As part of the programme of annual direct ground counts, we made direct counts on foot in February 2007 and 2008 to estimate elephant abundance. The team followed the methods and used the same transects prescribed for the ranch's regular annual counts: 34 north-south transects that spanned the whole study area. The line transect survey method was used (Buckland, Anderson, Burnham, & Laake, 1993; Buckland, Anderson, Burnham, Laake, Borchers, & Thomas, 2001). The survey team consisted of three people who walked in a straight line on a pre-determined compass bearing. The team walked slowly and stopped each time an elephant group was seen. The number of elephants, the group's compass bearing and the distance from the transect to the center of the group was recorded. Later the perpendicular distance of the group from the transect centre-line was calculated. The DISTANCE 4.1.2 software package was used to estimate the number of elephants from the perpendicular distance and group size data (Thomas et al., 2003, 2010).



Figure 1. Map of Burkina Faso showing the location of Nazinga Game Ranch

3.3 Dung Counts

The dung-pile population is related to the animal population by the following formula (Laing, Buckland, Burn, Lambie, & Amphlett, 2003):

$$\widehat{D}_a = \widehat{D}_s / (\hat{p}.\hat{t}) \tag{1}$$

Where \hat{D}_a is the estimated density of animal signs in the study area, \hat{D}_s is the estimated animal's dung-pile density; \hat{t} is the estimated mean time to decay of the animal signs present when the survey to estimate sign density is conducted, and \hat{p} is the estimated rate of production of signs per animal during the period preceding

the survey.

For a given animal population of a specific area, \hat{t} and \hat{p} are influenced by many environmental factors including seasonal variations (rainfall, temperature, termites activities), diet, animals' physiology and ages, size of dung, etc. However, if successive surveys are conducted at the same time of year, that is, \hat{t} and \hat{p} are constant from year to year, then equation 1 could be written as:

$$\widehat{D}_a = c.\,\widehat{D}_s\tag{2}$$

Where c is a constant $= 1/(\hat{p}.\hat{t})$.

According to this model, the trend in dung-pile density \hat{D}_s will reflect changes in the animal population's density \hat{D}_a .

In a tropical site with one wet season and one dry season each year, all dung-piles decay rapidly during the wet season as a consequence of the heavy rainstorms and high temperatures (Jachmann & Bell, 1979, 1984). During the dry season dung-piles dry out quickly and the decay rate is minimal and so dung-piles accumulate during the dry season (Jachmann & Bell, 1979, 1984). Our dung counts were conducted in April-May (the peak of the dry season) in 2007 and 2008, so \hat{D}_s is an estimate of the density of accumulated dung during the dry season.

We used the method now widely adopted for evaluating the abundance of elephant dung (Hedges & Lawson, 2006). The transect design differed from the direct counts. A grid of sides 2 km was placed over the study area. Then from a random start 54 transects were laid at 4 km intervals, such that each ran through the center of a selected cell (i.e, every other cell). Each transect was 1 km long and oriented north-south or east-west so as to cut across the drainage lines (Figure 1). The spacing of the transects ensured that they were not spatially correlated (Hema, Barnes, & Guenda, 2010). This design was used for both dung counts.

The line transect survey method was used to estimate dung pile abundance (Barnes, 1993; Buckland et al., 1993). The survey team consisted of three people who walked in a straight line on a pre-determined compass bearing. The following notes were made for each dung-pile: distance along transect, the stage of decomposition and the perpendicular distance, measured with a tape, from the transect centre-line. Only those dung-piles that fell in the decomposition stages A to D defined by Barnes and Jensen (1987) were included in the analysis. Estimates of dung-pile abundance were obtained using the DISTANCE 4.1.2 software package (Thomas et al., 2003, 2010).

3.4 Comparison of Estimates

Estimates were compared using Norton-Griffiths (1978) d-test:

$$d = (Y_2 - Y_1) / \sqrt{(\text{var}Y_2 + \text{var}Y_1)}$$
(3)

Where Y_1 and Y_2 were the first and second estimates respectively, and $var Y_1$ and $var Y_2$ were their variances.

3.5 Power Calculation

We used the free software TRENDS (Gerrodette, 1993) to calculate the power of each method to detect a trend. This software allows the user to choose one out of several options for each parameter. We set an exponential decline of 10% per annum. We chose an exponential decline, rather than the type whereby the population decreases by the same number each year, because Caro (2008) fitted straight lines to log plots of population size, indicating exponential declines, for a wide range of large mammals. We set the CV to be that estimated for the 2007 counts. The CV was set to be inversely proportional to \sqrt{N} (Gerrodette, 1987), since with small elephant populations the variance tends to increase as populations decline (Barnes, 2002). The test was one-tailed, since a manager usually has anecdotal information indicating which way his population is trending. We set α (the probability of a Type I error) to 0.05.

4. Results

4.1 Elephant Counts 1982–2008

Since 1982 the estimates of elephant numbers show considerable variation, even between estimates using the same methods in the same year, as in 1985 (Figure 2).

Only 21 elephant groups were seen in the 2004 direct count, and only 22 in the 2006 count; these sightings were too few to justify calculating an estimate and are not shown in Figure 2.



Figure 2. Estimates of elephant numbers in Nazinga Game Ranch from all types of survey from 1982 to 2008

Confidence intervals have been omitted for clarity. Additional data were collected from C. Lungren (pers. comm.); Bousquet (1982) in Jachmann (1988); O'Donoghue (1985); Jachmann (1988, 1991); Damiba and Ables (1994); Cornelis (2000); Ouédraogo et al. (2009); Bouché, Lungren, and Hien (2004); Hema, Ouédraogo, Belemsobgo, and Guenda (2007).

4.2 Direct Ground Counts in 2007 and 2008

The survey teams covered a total transect length of 680.2 km for each count. Forty-six elephant groups were seen in 2007 (Table 1). The median group size was 5 (mean = 7.4). The elephant population estimate was 2518 (95% confidence interval: 1476–4294) or 2.57 elephants.km⁻² (95% confidence interval: 1.51–4.38). This figure was considerably higher than any preceding estimate (Figure 2). The CV was 27.5%.

The histogram of perpendicular distances for 2007 shows that fewer elephants were seen close to the transect's centre-line than in the 25-49 metre class (Figure 3).

The following year only 27 elephant groups were seen on the same transects (Table 1). The median group size was 4 (mean = 5.9). The elephant population estimate was 1134 (95% confidence interval: 503-2553) or 1.16 elephants.km⁻² (95% confidence interval: 0.51-2.60). The CV was 42.0%.

While the 2008 count represented an apparent decrease of 55% compared to 2007, the difference was not significant (d = -1.64, NS).

Comparison of the earlier direct counts and the most recent direct counts suggests an increase (Figure 2): the three counts in 1985 and 1987 gave a median estimate of 487 elephants compared to a median of 1363 elephants for the three most recent ones in 2005, 2007 and 2008.

Parameter	Elephant	Elephant population
	population	2008
	2007	
Number of observations (<i>n</i>)	46	27
Estimator	Uniform/Cosine	Half-normal/cosine
Number of parameters used in the estimate	1	1
f(O)	0.010 227	0.009 860 3
var[<i>f</i> (<i>O</i>)]	1.42×10^{-6}	4.32x10 ⁻⁶
Effective strip width w (in metre)	97.78	101.42
Densityestimate per $\text{km}^2(D)$	2.57	1.16
Var(D)	0.50	0.24
95% Upper confidence limit	4.38	2.60
95% Lower confidence limit	1.51	0.51
% CV	27.51	42.04
Population estimate (N)	2518	1134
95% Upper confidence limit	4294	2553
95% Lower confidence limit	1476	503
X^2	4.76	2.24
p	0.783	0.134
df	8	1
Range of observed group sizes	1-27	1-18
Density of groups estimate (in km ²)	0.35	0.20
Mean group size estimate $(E(S))$	7.42	5.93

Table 1. Estimates of the size of the elephant population in two consecutive years in Nazinga Game ranch. The data were collected by direct observations on the ground using the line transect method (Buckland et al., 1993; Thomas et al., 2003, 2010)

NB: The 2008 data were truncated at 140 metre.



Figure 3. Histogram of perpendicular distances for the direct ground count of elephants in 2007. The fitted probability density curve is also shown

4.3 Dung Counts

Each dung count covered a total transect length of 54 km. In 2007 the team recorded 2579 dung-piles (Table 2). The dung-pile population estimate was 1,828,800 (95% confidence interval: 1,381,000–2,421,900) or 1864.3 dungpiles.km⁻² (95% confidence interval: 1407.7–2468.8). The CV was 14.1%. The histogram of perpendicular distances (Figure 4) showed a good shoulder at the centre-line and a smooth detection curve.

In 2008 the team recorded 3819 dung-piles on the same transects (Table 2). The dung-pile population estimate was 2,230,400 (95% confidence interval: 1,537,800–3,234,800) or 2274 dungpiles.km⁻² (95% confidence interval: 1568–3298). The CV was 18.7%. The observed frequencies and the fitted detection curve were similar to the previous year, but the higher χ^2 value indicated that the fit was not as good (Table 2).

While the 2008 count represented a 22% increase compared to 2007, in contrast to the direct counts which suggested a decrease, the difference was not significant (d = 0.82, NS).

Table 2.	Estimates	of elephant	dung abun	dance in tw	o consecuti	ve years	in Nazinga	a Game ranch	. The da	ita were
collecte	d using the	e line transec	t method (E	Buckland et	al., 1993; 7	homas et	t al., 2003,	2010)		

Parameter	Dung population	Dung population
	2007	2008
Number of observations (<i>n</i>)	2579	3819
Estimator	Half-normal/Cosine	Half-normal/Cosine
Number of parameters used in the estimate	4	3
f(O)	0.07881	0.06490
var[<i>f</i> (<i>O</i>)]	0.33551 x 10 ⁻⁵	0.12522 x 10 ⁻⁵
Effective strip width w (in metre)	12.69	15.41
Densityestimate per $\text{km}^2(D)$	1864.30	2273.6
Var(D)	69000.78	180 854.57
95% Upper confidence limit	2468.8	3297.5
95% Lower confidence limit	1407.7	1567.6
% CV	14.09	18.70
Population estimate (N)	1 828 800	2 230 400
95% Upper confidence limit	2 421 900	3 234 800
95% Lower confidence limit	1 381 000	1 537 800
X^2	18.73	67.64
р	0.539	0.000
df	20	21



Figure 4. Histogram of perpendicular distances for the elephant dung count; the fitted probability density curve is also shown

4.4 Power Calculation

With a 10% annual decrease in elephants, after 5 years direct counts would have a 22% probability of detecting a significant change in population size (Figure 5). In contrast, dung counts would have a 50% probability of detecting a significant change after 5 years. Direct ground counts would require 10 years of consecutive counts to have 80% power to detect a significant change (by convention 80% power is usually taken as the standard) whereas dung counts would exceed 80% by year 7.



Figure 5. Comparison of power curves for direct and dung counts (each curve shows the power to detect an exponential decrease in numbers with a series of annual counts)

5. Discussion

The dung counts were more precise, since they had smaller CVs (Norton-Griffith, 1978; Buckland et al., 1993, 2001), than the direct counts. Therefore there was a higher probability of detecting a change in numbers using dung counts. In other words, greater statistical power was achieved with dung counts. Furthermore, the dung counts were far more efficient, since the smaller CVs were obtained with transects that totaled 54 km compared to 680.2 km for the direct ground counts. From the perspective of monitoring trends in elephant populations, dung counts are superior to direct ground counts.

However the dung count method rests upon the assumption that c (in Equation 2) is constant from year to year. Dung decay rates vary with season and habitat type (Barnes, 1982; Laing et al., 2003). Rainfall affects the quality of the vegetation, and that affects the digestive processes and therefore the rate of defecation (Barnes, 1982). Rainfall also accelerates the rate of dung decay (Barnes, Asamoah-Boateng, Naada-Majam, & Agyei-Ohemeng, 1997). Thus rainfall is the most important factor to affect the value of c. During the two years there were no evident change in habitat and so dung production rates were assumed constant. By executing our surveys at the same point in each year (in April, the late dry season) we assume that we eliminated any significant variation in dung decay rate. However, this assumption must be tested. An alternative hypothesis, which should also be tested, is that at the peak of the dry season c is a function of accumulated temperature and accumulated rainfall since a fixed date, either 1st January or the end of the previous wet season.

The variation in Nazinga elephant population estimates over the years is partly due to the movements of elephants in and out of the area which is linked by corridors to the Kaboré-Tambi NP (formerly Po NP) and the Nazinon valley to the northwest, and to the Sissili valley and Ghana to the south. It is also a consequence of the plethora of methods that has obscured any real trends. Consequently there is not a series of accurate estimates using the same method from which the trend can be estimated. The direct counts on the ground were planned to standardize the annual counts so as to provide an objective measure of trend from 1982. However, this method is

inefficient and inaccurate, and it gives inconsistent results. For example, in 2007 the direct count estimated 2518 elephants for the RGN, the highest estimate ever recorded. The following year the same team walking exactly the same transects estimated less than half that number. In both years the counts failed to achieve the minimum number of observations required for a reasonable estimate (Buckland et al., 1993; Jachmann, 2001), despite covering 680 km of transect in each year.

Furthermore, in both years fewer elephants were seen close to the centre-line than in the 25-49 metre class (Figure 3). This contravenes one of the most important assumptions of line transect method: that all animals on the transect center line are observed with certainty (Buckland et al., 1993, 2001). This problem is due to elephants detecting the approach of the surveyors and moving away from the line of march before they were seen. The consequence of both the small number of sightings and the lack of a clear shoulder for the detection curve is an estimate lacking both accuracy and precision.

There is an additional problem common to line transect surveys of animals: the perpendicular distance was not measured directly. Rather, the animal-to-observer distance (or radial distance) was measured and then the perpendicular distance was later calculated by simple geometry. The detection function for the animal-to-observer distance gives a biased estimate of the probability density function on which the distance estimate depends (Buckland, Plumptre, Thomas, & Rexstad, 2010). Therefore any estimate of elephant abundance produced by these direct counts is likely to be inaccurate (Buckland et al., 2010).

The direct ground counts required a huge sampling effort but gave a poor return: 0.05 groups recorded per km of transect for the 2007 and 2008 counts combined. In contrast, the encounter rate for the dung counts was orders of magnitude higher: 59.24 dung-piles per km of transect for the two counts combined. The histogram of perpendicular distances (Figure 4) is an excellent example of a visibility curve showing how the probability of detection decreases with distance from the transect center line. Consequently the estimate of dung-pile abundance is likely to be accurate.

The direct counts provided an instantaneous picture of the distribution of elephants across the study area. Consequently there was great variation between transects, with most transects recording no elephants at all. In addition, the elephant groups varied in size. These factors increased the variance of the direct estimate. In contrast, dung is more evenly distributed than an instantaneous snapshot of the animals, and the variation between transects is low (Jachmann, 1991). Consequently the CVs of the dung counts (14.1% and 18.7% in 2007 and 2008 respectively) were considerably smaller than those of the direct counts (27.5% and 42.0% respectively). Therefore, the dung counts were considerably more precise and conferred greater statistical power to detect changes in elephant abundance.

Ouédraogo, Delvingt, Doucet, Vermeulen, and Bouché (2009) also recognized the problems of direct ground counts and tested a total ground count of elephants in the same Nazinga study area. Blocks were searched while walking along reconnaissance transects and signs of elephants were followed up until the elephants were sighted. This method may work well when there is only a small number of elephant groups and the study area is small, but is likely to become increasing inaccurate with larger populations and with an extensive area to cover (Ouédraogo et al., 2009). Total ground counts have also been used in India but need to be carefully controlled to avoid double-counting some groups and missing others (Lahiri-Choundhury, 1991). Whereas total counts from the air may give accurate estimates, total counts by observers on foot are open to too many potential biases for regular monitoring programmes.

Dung counts also have potential sources of errors such as observer efficiency, difficulty of distinguishing whether adjacent dung-piles represent a single defaecation or two or more defaecations, and determining whether a dung-pile has decayed to the point where it should not be recorded (Barnes, 1993; Hedges & Lawson, 2006). Transforming dung density into elephant density results in wider confidence limits because the standard error (SE) of defaecation and of decay each contribute to the standard error (SE) of elephants numbers (Barnes, 1993). But this is an unnecessary complication if one is concerned only with evaluating trends in the elephant population. Here we advocate using dung density as the index of elephant abundance and an index of population change. There is no need to go through the costly exercise of estimating dung decay rates and defaecation rates, because it is not necessary to convert dung density into animal density when monitoring the same area year after year. As long as dung counts are conducted at the same time of year so that conditions are standardized–for example, the end of the dry season–then estimates of dung density will provide valid estimates of trend.

This raises the fundamental question: what is the purpose of counting animals in a protected area? A manager may like to know how many individuals of a particular species are in his domain. But in most cases the trend is more important than the actual number, and to evaluate the trend one needs precise estimates. For this purpose an

index of population size that is precise is more useful than an estimate of animal numbers (Jachmann, 2001). Direct ground counts provide estimates with lower precision that dung count. Hence, dung counts are most likely to give the required information on trends. Furthermore, aerial sample counts are also unlikely to give useful data on trends, especially in the West African context where dung counts give estimates that are more precise than those from aerial sample surveys (Barnes, 2002). Even when animals are at low density, there will always be plenty of dung in the dry season because the rate of dung decay is zero while animals will continue to defecate (Jachmann & Bell, 1979, 1984).

We recommend that for monitoring elephant trends at Nazinga, a dung count be conducted at the same time each year: the peak of the dry season. Once one has established the transect design, the same transects can be sampled year after year. The methods of data collection are simple and wildlife staff can be easily trained to do this work. The dung counting methods outlined in this paper are simple and practical, and will allow the wildlife manager to monitor the abundance and distribution of all large herbivore populations, not just elephants, year after year.

With growing concern about the decline of wildlife populations in protected areas (Brashares, Arcese, & Sam, 2001; Craigie et al., 2010), it will become increasingly important to evaluate trends. As populations diminish, the relative advantages of dung counts over other methods will increase (Barnes, 2002). Therefore we recommend that wildlife biologists should increasingly immerse themselves in dung. We recommend that monitoring programmes follow the protocol developed by Hedges and Lawson (2006) for dung counts. These arguments apply to elephants but may well be applied to all herbivores at Nazinga.

Resources for wildlife studies are scarce yet the needs to evaluate trends are urgent because of expanding human pressures around protected areas like Nazinga (Hema et al., 2010). Dung counts can be applied to all savannas, bushlands and woodlands elsewhere across the continent. They may apply not just to protected areas of modest size but also to large ecosystems where estimates of trends for all herbivores are needed (Caro, 2008). Furthermore, where poaching has been intense, and animals are shy, dung counts will give an estimate of trend better than the direct ground count. However, many wildlife managers are skeptical about the validity of dung counts in general. We suggest that the validity of dung counts for monitoring large herbivores in open country should be verified at a large site that already has a well-established monitoring program.

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